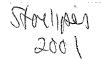
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The Effect of Depuration on Transmission of Aeromonas salmonicida between the Freshwater Bivalve Amblema plicata and Arctic Char

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Abstract.—A model system was used to study bacterial fish pathogen transmission between the freshwater bivalve Amblema plicata and two strains (Nauyuk and Labrador) of Arctic char Salvelinus alpinus. Aeromonas salmonicida, the cause of fish furunculosis, was readily transmitted from Arctic char to A. plicata and vice versa via simple cohabitation. Clinical furunculosis was artificially established in Nauyuk Arctic char via horizontal exposure to Labrador Arctic char that received intraperitoneal injections of A. salmonicida. After the Nauyuk Arctic char began to die, A. plicata were placed in the tank with the fish. After 33 d of cohabitation, a group of 10 A. plicata was cultured, and A. salmonicida was isolated from all 10. The remaining A. plicata were transferred to other tanks being supplied with specific-pathogenfree water. At 1, 5, 15, and 30 d after transfer, 60 uninfected Labrador Arctic char were cohabitated with the A. plicata. Transmission of A. salmonicida from A. plicata to the Arctic char was evaluated via fish mortality and bacterial culture after 3-4 weeks of exposure. Mortality to A. salmonicida occurred in groups exposed to A. plicata after 1 and 5 d of depuration but not in groups exposed after 15 and 30 d. The bacterium was not isolated from either the A. plicata or the Arctic char in the 15- and 30-d groups. Results indicate that the current minimum 30-d quarantine of freshwater bivalves destined for relocation to prevent spread of zebra mussels Dreissena polymorpha is sufficient to allow depuration of a fish pathogen and, thus, to prevent the spread of disease.

There are approximately 300 species and subspecies of freshwater bivalves that are native to the waters of North America. Williams et al. (1993) reported that greater than 70% of the species and subspecies are categorized as endangered, threatened, or of special concern. The survival of freshwater bivalves depends on the quality of their environment. Furthermore, they utilize fish in part of their life cycle for development of glochidia; thus adverse effects placed on the intermediate fish hosts directly impact development of the bivalve larvae and subsequent numbers of juvenile ani-

mals. Examples of adverse effects include siltation from runoff due to agriculture or construction (Ellis 1936; Kat 1982) and waterway construction or activities such as dredging and dam building (Fuller 1974; Keller and Zam 1990). The combined effects of these activities with the inadvertent introduction of zebra mussels *Dreissena polymorpha* have caused concern for the continued survival of large-river native species (Hebert et al. 1991; Nalepa 1994).

The U.S. Fish and Wildlife Service along with other federal and state agencies initiated efforts in the mid-1990s aimed at conserving native freshwater bivalves in selected large-river systems despite the advancing spread of zebra mussels. One such effort was to collect native animals and transfer them to safe locations. The goal is to maintain and propagate them and to reintroduce them at an appropriate future date when impacts caused by zebra mussels are of lesser concern. Sites to maintain the native bivalves include salmonid-rearing hatcheries.

The concerns for preventing and the consequences of the transmission of fish pathogens via transfer of fish or fish eggs between facilities are well documented (Piper et al. 1982). This is particularly relevant to hatcheries because they typically rear fish under intensive culture, which may render the fish more susceptible to diseases (Wedemeyer 1996; 1997). Fish health policies include guidelines and restrictions based on the premise of preventing pathogen transfer to naïve facilities and fish. The primary means for prevention is via fish health inspections and assigned health certifications. The potential for transfer of fish pathogens also is relevant to relocation of freshwater bivalves from natural river waters to hatcheries. Bivalves, being filter feeders, could be exposed to fish pathogens in their environment and acquire them via filtering. The potential for pathogen transfer with the relocated bivalves is not known, nor is an effective time that is needed for bivalve

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depuration of infectious agents. This study was done in an attempt to address these questions.

A model system was used to evaluate the transmission of pathogens between fish and bivalves. In a study by Starliper and Morrison (2000), groups of bivalves were exposed by waterborne challenge to various fish pathogenic and nonpathogenic bacteria previously isolated from fish. No mortality or other obvious adverse effects to the bivalves were noted after the challenge. However, one group that was exposed to Aeromonas salmonicida, the cause of furunculosis, followed by a 24-h period of depuration and subsequent cohabitation with a susceptible fish host resulted in disease and death to the fish. From this it was determined that this experimental situation could be used as a model to study the potential for pathogen transmission.

Methods

Amblema plicata were collected by brailing during the June 1999-collecting season from the Ohio River (river mile 175), near Muskingum Island and adjacent to Wood County, West Virginia. This region is not easily accessible for public use, and the mussel beds are protected by the Ohio River Islands National Wildlife Refuge. Harvesting and collection are only allowed by permit for scientific and management purposes. The A. plicata were quarantined for 60 d at an enclosed holding facility. Zebra mussel veligers were detected after the first 30 d; thus the A. plicata spent another 30 d in quarantine, at which time they were determined to be free of zebra mussels (Chaffee 1997). After quarantine, the A. plicata were transported by live well (truck) to the Leetown Science Center where they were placed in 566-L aquaria (density = 21.4 kg/m³) supplied with 24°C reservoir water until used for experimental purposes. A substrate (20 cm deep) consisting of sand and small gravel covered the bottom of each aquarium; dissolved oxygen was between 8 and 9 mg/L, and the water flow was about 30 L/min. The A. plicata received food via the water that was comprised primarily of chlorophytes and diatoms at a density ranging from 2.00×10^3 to 2.90×10^5 cells/mL (R. Villella, U.S. Geological Survey, personal communication).

Two strains of Arctic char Salvelinus alpinus were used, provided by the Conservation Fund's Freshwater Institute, Shepherdstown, West Virginia. The Labrador strain averaged 250 g each, and the Nauyuk strain averaged about 500 g each. The fish were originally received as eggs from brood-

stock that were certified as specific-pathogen free (Office of the Federal Register 1999). The fish were held indoors, in 1,130-L circular tanks at a density of less than or equal to 70 kg/m³ that were fed with flow-through spring water at 40 L/min (12°C; pH 7.0–7.2; dissolved oxygen was 10–12 mg/L). The artificial lighting was adjusted to conform to the natural day/night cycle. The fish were fed 1.5% of their body weight per day with a 38% protein diet (Aquaculture Food; Zeigler Brothers, Inc., Gardners, Pennsylvania).

The strain of A. salmonicida (AS-98-1) was originally isolated from a population of Atlantic salmon Salmo salar used for research at the Leetown Science Center in 1998. The isolate conformed biochemically to criteria typical of the species (Hiney and Olivier 1999). The bacteria were grown in 100 mL of tryptic soy (TS) broth (Difco Laboratories, Detroit, Michigan) at 20°C on a 100-rpm shaker for 48 h. Tenfold dilutions of the cells were prepared in TS broth for challenge and for determination of viable cell numbers. Drops (0.025 mL) of each dilution were placed on the surface of TS agar with 0.01% Coomassie brilliant blue (CBB medium; Bio-Rad Laboratories, Hercules, California) (Cipriano and Bertolini 1988). The resulting bacterial colonies were enumerated, and the number was converted to colony-forming units (CFU) injected per fish.

Isolation of A. salmonicida from fish was done by inoculation of kidney and mucus onto CBB plates by the use of one of two methods. One was a viable cell quantification technique; tissues were collected aseptically and placed in preweighed tubes, the tubes were reweighed, and the tissues were diluted 1:10 and homogenized in sterile 0.1% peptone and 0.05% yeast extract (pep-ye; Difco). Additional 10-fold dilutions were prepared for each sample in pep-ye and used to inoculate CBB as previously described. Mucus was collected as described by Cipriano et al. (1992). Numbers of A. salmonicida were reported as CFU per gram of kidney or mucus. Cell counts of total bacteria were also determined on the same CBB plates; this equaled A. salmonicida plus all other bacteria that grew. Streak plating was the other technique. Two plates (to enhance isolation probability) were inoculated for each tissue and from each fish. All plates were incubated at 20°C for 48 h. Presumptive, blue colonies were transferred to fresh medium for biochemical characterization of A. salmonicida by the use of standard identification methods (MacFaddin 1980; Koneman et al. 1983; Hiney and Olivier 1999).

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Bacteria were isolated from A. plicata by use of the techniques of Starliper et al. (1998) with modifications. Physical data (weight, length, and width) were recorded from each animal. Before opening the valves, the external surface of the shells was gently scrubbed clean by use of a brush and 200 mg of sodium hypochlorite/L and then allowed to air dry. Aseptic technique was used for removing the internal soft tissues. The shells were pried apart, the abductor muscles were cut, and the liquid inside of the shell but outside of the soft tissues (termed "fluid") was collected and quantified. Tenfold dilutions were then prepared in pep-ye, and 0.025-mL drops were placed on CBB plates. The soft tissues were removed and portioned as two samples. One sample, labeled "Gut," contained primarily digestive tract, and the other, labeled "OT," comprised the remaining soft tissues. To evaluate the presence of A. salmonicida in the soft tissues without influence from residual bacteria on their external surfaces or fluid, each soft tissue sample was surface disinfected by dipping in 200 mg of sodium hypochlorite/L for 30 s followed by rinsing in sterile pep-ye. Tissues were placed in preweighed laboratory bags (Tekmar, Cincinnati, Ohio); the tissues were weighed and diluted 1:2 (w/v) in pep-ye and homogenized for 2 min in a stomacher (model 80 Laboratory Blender; Seward Medical, London, United Kingdom). About 2 mL of the resulting liquid homogenate was transferred to a sterile tube for ease of handling, and from this, a series of three 10-fold dilutions were prepared in pep-ye. Drops (0.025 mL) from each dilution were placed onto the surface of CBB plates. The plates inoculated from the fluid, Gut, and OT dilutions were incubated at 20°C for 48 h. Presumptive and blue A. salmonicida were enumerated and picked for biochemical characterization to confirm their identity. Total bacterial colony numbers were also determined.

The presence of A. salmonicida in the tank was determined by collecting 1 L of water effluent; this was centrifuged for 30 min at 4°C at 5,000 × gravity (Sorvall RC2-B, HS-4 rotor). The supernatant was discarded, and the pellet was suspended in 5 mL of pep-ye. A series of 10-fold dilutions was prepared, and 0.025-mL drops were placed on the surface of CBB plates. Aeromonas salmonicida and total bacteria were reported as CFU per milliliter of tank water. The blue colonies were confirmed as A. salmonicida as described previously.

An artificially induced clinical case of furunculosis in Nauyuk Arctic char served as the source of *A. salmonicida* for the *A. plicata*. Two groups

of 60 Labrador Arctic char were exposed to 2.26 \times 10² and 2.26 \times 10⁰ CFU/fish, respectively, by intraperitoneal injection. These fish were placed in a 900-L tank along with about 300 Nauyuk char. The challenge doses to the Labrador char were selected on the basis of a previous determination of a dose lethal to 50% of the fish (Reed and Muench 1938) of 5.12×10^{0} CFU/fish for the isolate used. Susceptibilities of both Arctic char strains to A. salmonicida were similar (data not presented). The two sizes of Arctic char were used to distinguish injected fish from those that were not. Cell concentrations were chosen to produce mortality in the Labrador char at a level sufficient to provide a horizontal exposure with A. salmonicida and subsequent development of furunculosis in the Nauyuk char. Amblema plicata (N = 127)were placed in the bottom of the fish tank when the first Nauyuk char died due to A. salmonicida. By this time, nearly all of the injected Labradors had died.

Before adding the A. plicata to the tank, 10 were sampled to determine their baseline levels of bacteria and to ensure A. salmonicida was not present by bacterial culture of Gut, OT, and fluid samples as previously described. The A. plicata were transferred to their respective clean tanks after a desired prevalence of A. salmonicida of greater than 50% was achieved. Prevalence was determined by the use of 10 A. plicata examined after 17 and 33 d of cohabitation. The remaining A. plicata (n = 107)were transferred to four previously disinfected tanks (152 L). Sixty uninfected Labrador char were added to each tank on days 1, 5, 15, and 30 after the A. plicata were introduced. Immediately before adding the char, depuration of A. salmonicida was evaluated by bacterial culture of tank water effluent. Also, 10 and 7 A. plicata from the 1- and 5-d groups, respectively, were cultured for A. salmonicida as previously described.

All transmission studies of A. salmonicida were conducted in tanks fed with spring water as previously described. The 900-L tank received 31.1 L/min, which equaled 2.1 turnovers/h. The four 152-L tanks were regulated to an equal rate of 5.83 L/min or about 2.3 turnovers/h. Amblema plicata were allowed to acclimate over a period of 2 d from reservoir water to spring water before co-habitation with the Nauyuk char infected with A. salmonicida.

Transmission of A. salmonicida from A. plicata to the Labrador char was presumptively determined via development of disease and mortality and confirmed by biochemical characterization of

TABLE 1.—Prevalence and viable cell counts of *Aeromonas salmonicida* (AS) from *Amblema plicata* after 17 and 33 d cohabitation with the diseased Nauyuk Arctic char and following relocation after 1 and 5 d depuration in pathogen-free water. Also included are the mean viable cell counts for all (total) bacteria that grew on the CBB^a primary isolation plates. Abbreviations are as follows: OT = all other soft tissues, P:S = ratio of number of positive tissues to number sampled, CFU = colony-forming units.

Statistic	Gut ^b	OT	Fluid ^e
	Prevalence after	17 d	
AS P:S	1:10	2:10	3:10
AS CFU/g or mL	2.64×10^{4}	1.32×10^{5}	1.09×10^{3}
Range		2.64×10^{2} 2.64×10^{5}	5.20×10^{2} – 1.44×10^{3}
All bacteria, mean (CFU/g or mL)	1.18×10^{5}	9.61×10^{5}	2.91×10^{4}
	Prevalence after	33 d	
AS P:S	6:10	10:10	4:10
AS CFU/g or mL	4.20×10^{4}	1.20×10^{4}	2.50×10^{2}
Range	2.70×10^{1} – 1.60×10^{5}	1.00×10^{2} – 1.10×10^{5}	4.00×10^{1} -4.00×10^{2}
All bacteria, mean (CFU/g or mL)	1.60×10^{5}	6.40×10^{4}	4.63×10^{3}
	Prevalence after 1 d of	depuration	
AS P:S	3:10	1:10	1:10
AS CFU/g or mL	1.40×10^{2}	2.00×10^{2}	1.33×10^{2}
Range	2.00×10^{1} – 3.00×10^{2}		
All bacteria, mean (CFU/g or mL)	9.52×10^{4}	6.62×10^{4}	3.77×10^{4}
	Prevalence after 5 d of	depuration	
AS P:S	1:7	4:7	4:7
AS CFU/g or mL	1.00×10^{2}	2.78×10^{2}	3.67×10^{1}
Range		1.60×10^{3} – 1.00×10^{3}	$1.33 \times 10^{1} - 6.67 \times 10^{1}$
All bacteria, mean (CFU/g or mL)	6.96×10^{4}	1.10×10^{5}	2.64×10^{4}

^a Tryptic soy agar with 0.01% Coomassie brilliant blue.

the bacteria from the mucus and kidney of each fish. In the cohabitation groups with no mortality after 3-4 weeks, bacterial culture of *A. plicata* and char was done to determine the presence of *A. salmonicida*.

Results

The average length of each A. plicata was 100.7 mm, ranging from 57 to 128 mm. The width averaged 75.6 mm and ranged from 45 to 99 mm. The mean total weight was 243.6 g (range of 51.1–493.8 g); an average of 23.0 g (9.4%) of the total weight was comprised of soft tissues. Gut samples accounted for an average of 12.5 g (range of 1.2–24.1 g; 54.1%) of the total soft tissue, whereas OT samples comprised the balance of the total soft tissues and had a mean weight of 10.6 g and weights ranging from 2.3 to 18.4 g/sample. There was an average of 21.4 mL (3.5–47 mL) of pallial fluid collected from each A. plicata.

The first intraperitoneal-injected Labrador Arctic char died 7 d after challenge. The first Nauyuk Arctic char died 17 d after the initial challenge. The cause of death was confirmed biochemically as A. salmonicida. Regardless of the route of exposure, the counts of A. salmonicida from kidney

tissues were similar: 1.22×10^9 CFU/g for the injected Labradors and 1.10×10^9 CFU/g for the Nauyuk char. Total bacterial counts from the Labrador char kidneys were similar suggesting that A. salmonicida was the predominate bacterium in these fish. Likewise, total counts from the Nauyuk char were also similar to the counts of A. salmonicida (average, 1.12×10^9 CFU/g). The mean total bacterial counts from mucus were only slightly higher than the cell counts of A. salmonicida. However, higher numbers of A. salmonicida were isolated from the mucus of horizontally infected fish, 4.39×10^7 CFU/g compared with 1.71×10^7 CFU/g from the injected-Labrador mortality.

Aeromonas salmonicida was not isolated from the 10~A. plicata examined to ensure the pathogen was absent before cohabitation. Tissues and fluids from all A. plicata resulted in bacterial growth on CBB. Mean cell counts for Gut and OT were 1.84×10^4 and 1.88×10^5 CFU/g, respectively. The average number of bacteria isolated from the fluids was 2.04×10^4 CFU/mL.

After 17 d of cohabitation, A. salmonicida was cultured from 3 of 10 (30%) A. plicata, from 1 Gut sample, and from 2 OT samples (Table 1). The cell count for the positive Gut sample was 2.64×10^4

b Gut = primarily digestive tract tissues.

c Fluid = liquid collected from within the valves, but outside of the soft tissues.

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TABLE 2.—Prevalence and viable cell counts of 1-d and 5-d groups of Aeromonas salmonicida (AS) isolated from Amblema plicata after 23 and 26 d, respectively, of cohabitation with Labrador Arctic char. The A. plicata were the source of A. salmonicida to the Arctic char; after becoming infected, the Arctic char would then serve to reinfect the bivalves. Abbreviations are as follows: OT = all other soft tissues, P:S = ratio of number of positive tissues to number sampled, CFU = colony-forming units.

Statistic	Gut ^a	OT	Fluid ^b
		1-d group	
P:S	2:10	5:10	4:10
AS CFU/g or mL	6.25×10^{4}	2.48×10^{4}	3.55×10^{3}
Range	$6.10 \times 10^4 - 6.40 \times 10^4$	$1.00 \times 10^2 - 5.40 \times 10^4$	1.33×10^{1} – 1.07×10^{4}
		5-d group	
P:S	8:10	7:10	6:10
AS CFU/g or mL	3.09×10^{3}	8.39×10^{3}	1.40×10^{3}
Range	1.00×10^{1} – 2.20×10^{4}	1.00×10^{2} -4.80×10^{4}	1.33×10^{2} -4.00×10^{3}

^a Gut = primarily digestive tract tissues.

CFU/g. Counts for the positive OT samples were 2.64×10^2 and 2.64×10^5 CFU/g. The bacterium was isolated from three fluids at concentrations ranging from 5.20×10^2 to 1.44×10^3 CFU/mL. A second evaluation done on day 33 of 10 A. plicata revealed 100% A. salmonicida (10 OT, 6 Gut, and 4 pallial fluid; Table 1). Viable A. salmonicida from OT ranged from 1.00×10^2 to 1.10×10^5 CFU/g with a mean of 1.20×10^4 CFU/g. By day 33, mortality in the Nauyuk char had subsided to three to six fish per week, and total mortality in the population of fish was about 40%. At the time the A. plicata were transferred, kidney and mucous tissues of 60 fish were cultured on CBB for isolation of A. salmonicida. Kidneys from four fish were positive (mean of 6.73×10^6 CFU/g), and 55 of the mucous tissues were positive and had a mean of 5.87×10^5 CFU/g.

Results of depuration of A. salmonicida by A. plicata after 1 and 5 d are presented in Table 1. After 1 d of depuration, prevalence of the bacterium decreased from 6 to 3 of the 10 Gut samples (mean of 1.40×10^2 CFU/g) and from 10 to 1 of the OT and from 4 to 1 of the fluid samples. Also, 5.99 × 101 CFU of A. salmonicida/mL was isolated from the tank water effluent. Prevalence of A. salmonicida in the A. plicata was higher when evaluated after 5 d of depuration; positives were detected from four of seven OT samples and four of seven fluids, including two that were paired with the OT samples of the same bivalve. One Gut tissue was positive. However, A. salmonicida was not isolated from the tank water effluent. When the three samples from each animal were combined, A. salmonicida was isolated from 6 of 7 A. plicata after 5 d of depuration, whereas 3 of 10 were positive after 1 d of depuration. Aeromonas salmonicida was not isolated from tank water effluent from either of the groups after 15 or 30 d of depuration.

No A. salmonicida was detected in the Labrador char exposed to A. plicata 15 or 30 d after their removal from the Nauyuk char. In both the 1- and 5-d groups, disease and mortality to the fish resulted. The first dead fish in each group was recorded on the 14th day after the clean char were added. Aeromonas salmonicida was biochemically confirmed from kidneys and mucous samples from the dead fish in both the 1- and 5-d groups. The 1-d group was terminated 23 d after cohabitation was initiated. Within this time, mortality to the fish totaled 29 of 60 (48%). Aeromonas salmonicida was isolated from 8 of 31 kidneys and 11 of 21 mucous samples from the surviving fish. The 5-d group was terminated after 27 d; 34 of 60 (57%) char died, and A. salmonicida was isolated from 3 kidneys and 19 mucous samples of the fish that survived through 27 d. In both the 1- and 5-d groups, A. salmonicida increased in prevalence and concentration in the A. plicata (Table 2), relative to the levels in A. plicata evaluated after depuration alone at 1 and 5 d (Table 1).

The 15- and 30-d groups were terminated after 29 and 22 d, respectively, of cohabitation. No mortality to the Labrador char occurred in either of these groups. No A. salmonicida was isolated from the char or the A. plicata after conclusion of the observation periods.

Discussion

It is encouraging to fishery and bivalve resource managers that in the worst-case scenario, A. salmonicida was not experimentally transmitted to Arctic char at some time less than 15 d, which is

b Fluid = liquid collected from within the valves but outside of the soft tissues.

only one-half of the current minimum-recommended quarantine to prevent spread of zebra mussels. The model system used in this study provides strong evidence of the potential for A. salmonicida to be spread by Arctic char and the freshwater bivalve A. plicata. The A. plicata appears capable of acquiring and harboring A. salmonicida and disseminating viable cells such that disease and mortality occur to susceptible hosts. Aeromonas salmonicida provides an excellent bacterial pathogen to study transmission between bivalves and fish. The vectoring animal A. plicata readily acquired the pathogen from water, and the bacterium is highly pathogenic to many salmonid hosts.

In this study, what could be considered a worst-case scenario was established by using clinically diseased and dying Arctic char as the source of A. salmonicida for the A. plicata. Furthermore, transfer of A. plicata to clean tanks was not done until the prevalence of A. salmonicida reached 100%. Therefore, the A. plicata harbored a near-maximum CFU of A. salmonicida that could be anticipated by the use of infected fish as the source of the pathogen. Considering these experimental criteria, it would be very unlikely that bivalves would be exposed to this type of challenge of viable cells in large rivers because of the volume and dilution potential of the flowing water.

The number of A. plicata positive for A. salmonicida decreased from 100% to 30% after 24 h of depuration. This rapid depuration is also encouraging. There was a negligible change in the prevalence of A. salmonicida in A. plicata assayed at 5 d of depuration compared with the level in the 1-d group. Likewise, the mean number of CFU of A. salmonicida per gram was similar for the two groups. Of particular interest would be the mechanisms involved in depuration (i.e., displacement by other bacteria or active ingestion or phagocytosis).

Depuration does not appear to have a direct correlation with time after transfer away from the source of A. salmonicida. In a study by Starliper et al. (1998), the fish pathogen Flavobacterium columnare was isolated from a single A. plicata sampled from the Ohio River. Other bivalves collected at the same site and time were quarantined to allow depuration. When some of these were examined for bacteria after only 1 d of depuration, F. columnare was not isolated. These bivalves were naturally infected with the pathogen, and the prevalence was 7% compared with the 100% prevalence of A. salmonicida in the present study.

Depuration of A. salmonicida by A. plicata in

the present study contrasts with depuration of viruses by marine bivalves (Canzonier 1971; Mortensen et al. 1992). Mortensen et al. (1992) exposed Pecten maximus to infectious pancreatic necrosis virus (IPNV) by injection and bath challenge methods. Tissues from the infected scallops were examined for IPNV, and the virus was detected at least 11 months after challenge in injected animals and 50 d in the waterborne-exposed animals. Although IPNV titers only decreased after exposure, the authors surmised that the persistence of the virus may have been due to viral replication. The persistence of IPNV relative to that for A. salmonicida might be related to other factors such as the high challenge dose of virus, the bivalve species, or the methods of exposure. Also, the persistence could be explained by the virus being sequestered in digestive tissues and, thus, avoiding depuration (Canzonier 1971; Hay and Scotti 1986).

Canzonier (1971) exposed hard clams *Mercenaria mercenaria* to low levels of *Escherichia coli* and coliphage S-13. In 16° C seawater, the clams depurated *E. coli* from an initial concentration of 1.0×10^{5} CFU/mL of digestive tissue homogenate to less than 1 CFU/mL within 48 h. In contrast, clams were less efficient at phage depuration. After 6 d, the phage titer in digestive tissues was less than 1 log unit below the initial titer. Phage depuration by *M. mercenaria* was related to water temperature and was accelerated above 16° C (Canzonier 1971).

Future plans to initiate health inspections of freshwater bivalves in a manner similar to those done routinely on captive fish populations should follow mandatory quarantine procedures for the bivalve. In previous studies by Starliper et al. (1998), it was shown that the distribution of the bacterial flora of bivalves changes rapidly after a change in their aquatic environment and this included the ridding of the pathogen *F. columnare*. The quarantine would allow the bivalves an opportunity to depurate, and then an inspection for fish pathogens would likely yield a more accurate account of the bacteria that might be relocated to a hatchery concurrent with the relocated bivalves.

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References

- Canzonier, W. J. 1971. Accumulation and elimination of coliphage S-13 by the hard clam, *Mercenaria* mercenaria. Applied Microbiology 21:1024–1031.
- Chaffee, C. 1997. Ohio River valley ecosystem mollusk conservation plan: strategic action plan-phase II. Appendix H: draft freshwater mussel quarantine protocol. U.S. Fish and Wildlife Service, Bloomington, Indiana. 13 pp.
- Cipriano, R. C., and J. Bertolini. 1988. Selection for virulence in the fish pathogen Aeromonas salmonicida, using coomassie brilliant blue agar. Journal of Wildlife Diseases 24:672-678.
- Cipriano, R. C., L. A. Ford, J. D. Teska, and L. E. Hale. 1992. Detection of Aeromonas salmonicida in the mucus of salmonid fishes. Journal of Aquatic Animal Health 4:114-118.
- Ellis, M. M. 1936. Erosion silt as a factor in aquatic environments. Ecology 17:29-42.
- Fuller, S. L. H. 1974. Clams and mussels (Mollusca: Bivalvia). Pages 215–273 in C. W. Hart, Jr., and S. L. H. Fuller, editors. Pollution ecology of freshwater invertebrates. Academic Press, New York, New York.
- Hay, B., and P. Scotti. 1986. Evidence for intracellular adsorption of virus by the pacific oyster, Crassostrea gigas. New Zealand Journal of Marine and Freshwater Research 20:655-659.
- Hebert, P. D. N., C. C. Wilson, H. H. Murdoch, and R. Lazar, 1991. Demography and ecological impacts of the invading mollusk *Dreissena polymorpha*. Canadian Journal of Zoology. 69:405–409.
- Hiney, M., and G. Olivier. 1999. Furunculosis (Aeromonas salmonicida). Pages 341-426 in P. T. K. Woo and D. W. Bruno, editors. Fish diseases and disorders Volume 3, Viral, bacterial and fungal infections. CABI Publishing, New York, New York.
- Kat, P. W. 1982. Effects of population density and substratum type on growth and migration of *Elliptio* complanata (Bivalvia: Unionidae). Malacological Review 15:119-127.

- Keller, A. E., and S. G. Zam. 1990. Simplification of in vitro culture techniques for freshwater mussels. Environmental Toxicology and Chemistry 9:1291– 1296.
- Koneman, E. W., S. D. Allen, V. R. Dowell, Jr., and H. M. Sommers. 1983. Color atlas and textbook of diagnostic microbiology, 2nd edition. J. B. Lippincott Company, Philadelphia.
- MacFaddin, J. F. 1980. Biochemical tests for identification of medical bacteria, 2nd edition. Williams and Wilkins, Baltimore, Maryland.
- Mortensen, S. H., E. Bachere, G. L. Gall, and E. Mialhe. 1992. Persistence of infectious pancreatic necrosis virus (IPNV) in scallops *Pecten maximus*. Diseases of Aquatic Organisms 12:221–227.
- Nalepa, T. F. 1994. Decline of native unionid bivalves in Lake St. Clair after infestation by the zebra mussel, *Dreissena polymorpha*. Canadian Journal of Fisheries and Aquatic Sciences 51:2227–2233.
- Office of the Federal Register. 1999. Importation of live or dead fish, mollusks, and crustaceans, or their eggs. Title 50, Section 16.13. Wildlife and fisheries. Code of Federal Regulations, National Archives and Records Administration.
- Piper, R. G., I. B. McElwain, L. E. Orme, J. P. McCraren, L. G. Fowler, and J. R. Leonard. 1982. Fish hatchery management. U.S. Department of Interior, Fish and Wildlife Service. Washington, D.C.
- Reed, L. J., and H. Muench. 1938. A simple method of estimating fifty per cent endpoints. The American Journal of Hygiene 27:493-497.
- Starliper, C. E., and P. Morrison. 2000. Bacterial pathogen contagion studies among freshwater bivalves and salmonid fishes. Journal of Shellfish Research 19:251–258.
- Starliper, C. E., R. Villella, P. Morrison, and J. Mathias. 1998. Studies on the bacterial flora of native freshwater bivalves from the Ohio River. Biomedical Letters 58:85-95.
- Wedemeyer, G. A. 1996. Physiology of fish in intensive culture systems. Chapman and Hall, New York, New York
- Wedemeyer, G. A. 1997. Effects of rearing conditions on the health and physiological quality of fish in intensive culture. Pages 35–72 in G. K. Iwama, A. D. Pickering, J. P. Sumpter, and C. B. Schreck, editors. Fish stress and health in aquaculture. Society for experimental biology seminar series, no. 62. Cambridge University Press, New York, New York.
- Williams, J. D., M. L. Warren, Jr., K. S. Cummings, J. L. Harris, and R. J. Neves. 1993. Conservation status of freshwater mussels of the United States and Canada. Fisheries 18:6–22.